

1 **Bycatch mitigation of endangered marine life**

2 Mireia Villafáfila García¹, Antonio Carpio^{2,3*}, Marga L Rivas¹

3 ¹ Department of Biology, Institute of Marine Science INMAR, University of Cádiz, Spain.

4 ² Department of Zoology, Campus of Rabanales, University of Cordoba, 14071 Córdoba, Spain.

5 ³ Grupo de Sanidad y Biotecnología (SaBio), Instituto de Investigación en Recursos Cinegéticos,
6 IREC (UCLM-CSIC-JCCM), Ronda Toledo 12, 13071 Ciudad Real, Spain.

7

8 Corresponding author: Antonio J Carpio: a.carpio.camargo@gmail.com

9

10 **Author Contributions:** Conceptualization, A.J.C., M.L.R.: data cleansing and writing—
11 original draft preparation, M.V.; writing—review and supervision, A.J.C., M.L.R. All
12 authors have read and agreed to the published version of the manuscript.

13

14 **Acknowledgments:** M.V.G. is supported by the ‘Plan propio de estímulo y apoyo a la
15 Investigación y Transferencia. INICIA-INV: Iniciación a la investigación (2022-2023)’ of
16 the University of Cadiz. A.J.C. is supported by a ‘Juan de la Cierva’ contract (IJC2020-
17 042629-I) funded by MCIN/AEI/10.13039/501100011033 and by the European Union
18 Next GenerationEU/PRTR. M.L.R. was supported by Postdoctoral Research Contracts of
19 the Andalusian government (Spain) and FEDER EU funds. This work received financial
20 support from the Agencia Andaluza de Cooperación Internacional para el Desarrollo
21 AACID (projects EMEP 2022UI007_2022).

22

23

24

25

26

27 **ABSTRACT**

28 The fishing gear deployed by fishermen in seas and oceans throughout the world not
29 only captures target species but also unintentionally ensnares non-target species, a
30 phenomenon known as "by-catch". This unintended capture of marine life can represent
31 significant challenges for the fishing industry, with adverse impacts on both the
32 environment and species such as sea turtles, marine mammals, seabirds and
33 elasmobranchs, which may be injured or even killed. To address this problem, the fishing
34 industry has implemented regulations and mitigation measures. In this literature review,
35 we have examined 389 articles published between 2010 and 2022 that assess the
36 effectiveness of these measures. It has been demonstrated that the most effective
37 measures are 'pingers' for marine mammals, 'TEDs' (Turtle Excluder Devices) for sea
38 turtles, and 'BSLs' (Bird Scaring Lines), more commonly known as 'tori lines', for seabirds.
39 The most complex case is that of elasmobranchs, and the most effective measure has
40 yet to be discovered. This complexity arises from the ongoing targeted fishing of these
41 species, resulting in less monitoring of their catches and, therefore, fewer surveys.
42 Overall, we encourage the global implementation of these measures by the fishing
43 industry in order to reduce by-catch in an attempt to ensure the future of many
44 endangered species.

45 Keywords: Incidental fishing · cetaceans · LEDs · marine turtles · marine birds · rays · sea
46 turtles · sharks

47 **1. INTRODUCTION**

48 Fishing poses a significant threat to marine vertebrates, including sea turtles, marine
49 mammals, seabirds and elasmobranchs. This threat stems from the impact of the fishing
50 industry on these species, both as a direct target and incidentally. The underlying cause
51 of this issue is the extensive presence of fishing fleets on the various seas and oceans. In
52 2020, the global fishing fleet was estimated to consist of approximately 4.1 million
53 vessels, with Asia being the predominant continent, accounting for a total of 2.68 million
54 vessels (FAO, 2022). These vessels utilize a range of fishing gear, depending on their
55 target species. The most common gear types include gillnets, longlines, purse seines with
56 and without purse lines, trawl nets and traps (FAO, 2022).

57 This fishing gear has a substantial impact on oceans and their ecosystems, resulting in a
58 series of adverse effects. These include an overexploitation of species, such as
59 overfishing, which can lead to the decline and even collapse of populations, ultimately
60 affecting marine biodiversity and ecosystems (Crowder et al. 2008). Certain fishing gear
61 also destroys habitats, as occurs with that used for bottom trawling since it damages
62 coral reefs, seagrass beds and seabeds (Crowder et al. 2008). Furthermore, the
63 disruption of the food web caused by the decline of specific species as a result of
64 overfishing can affect other species dependent on them as a food source, leading to a
65 cascading effect known as the 'top-down' or 'bottom-up' effect (Crowder et al. 2008).
66 Additionally, non-target or by-catch species are often captured alongside the intended
67 target species when non-selective fishing methods are employed, and a wide variety of
68 species is captured, including those not intended for capture or trade (Crowder et al.
69 2008). This includes species that are commercially undesirable owing to their lower value
70 or failure to comply with size or weight regulations, along with protected or endangered

71 species (which may have serious implications for their conservation) and non-
72 commercial species that are not of interest for trade (Agardy, 2000). In some cases, these
73 non-target species are not only captured but also injured or even killed, and these
74 species include marine mammals, sea turtles, seabirds and elasmobranchs (Crowder et
75 al. 2008).

76 Certain management measures with which to mitigate the impacts on these species have
77 been implemented, including catch and net restrictions and the deployment of specific
78 fishing technologies tailored to the groups affected (Lucas and Berggren, 2023). In the
79 case of sea turtles, physical and visual techniques have been tested, such as the Turtle
80 Excluder Device (TED) (Warden, 2011) and Light Emitting Diodes (LEDs), which could also
81 be a good alternative as regards mitigating sea turtle by-catch. In the case of marine
82 mammals, acoustic, physical, visual and echolocation measures have also been studied,
83 and acoustic deterrent devices such as pingers have been used. However, some studies
84 have demonstrated that certain species may become habituated to pingers over time
85 (Moan and Bjørge, 2021). With regard to seabirds, olfactory, physical and visual
86 measures are used to mitigate by-catch, with Bird Scaring Lines (BSLs), more commonly
87 known as tori lines, being the visual measure that has been most widely used in longline
88 fisheries (Domingo et al. 2017). In the case of elasmobranchs, all types of measures have
89 been tried, i.e., acoustic, olfactory, physical, visual, echolocation and electrosensory.
90 However, the knowledge regarding their effectiveness is limited, as in the case of
91 employing tori lines to reduce shark by-catch (Seidu et al. 2022; Jiménez et al. 2019).

92 In order to obtain a global overview, the principal objective of this study is to conduct a
93 comprehensive review of the existing scientific literature focused on assessing the

94 effectiveness of specific mitigation measures as regards reducing the by-catch of
95 endangered species groups, including sea turtles, marine mammals, seabirds and
96 elasmobranchs. As secondary objectives, this study aims to: (i) identify the mitigation
97 measures most commonly employed to minimize by-catch on a global scale; (ii) analyze
98 the geographic distribution of by-catch in fisheries by country in order to evaluate
99 variations, and (iii) provide an overview of the worldwide mitigation measurements that
100 have proven to have the greatest effectiveness as regards reducing by-catch for each
101 group. The overall objective of this research is to contribute with valuable insights into
102 mitigating the effect of by-catch on endangered marine species and to shed light on
103 measures that have proven to be the most successful, along with areas in which further
104 research is required.

105 **2. MATERIALS AND METHODS**

106 **2.1. Literature review**

107 This review was carried out using the Scopus, Web of Science and Google Scholar search
108 engines to search for publications containing citation indices spanning from 2010 to
109 2022. The keywords used were a combination of “bycatch” AND “fisheries” AND
110 “mitigation” AND “measures” AND “sea turtles” or “marine mammals” or “seabirds” or
111 “elasmobranchs”. These four groups of marine vertebrates were selected as the
112 objective of this study. The data collected were classified according to the animal class,
113 species name in English, scientific name, UICN category, the mitigation measure used,
114 the percentage of reduction in by-catch achieved, and finally, the reference of each
115 paper (Table 1 SI).

116 2.2. Species and areas

117 We collected global data and information from scientific literature concerning the by-
118 catch of the four marine vertebrate groups, such as sea turtles, marine mammals,
119 seabirds and elasmobranchs, by fisheries in the different seas and oceans worldwide.
120 Species of sea turtles included in the review belonged to the Cheloniidae and
121 Dermochelyidae families. These species are distributed in seas and oceans throughout
122 the world (except those in the Arctic and Antarctic), with the exception of Kemp's ridley
123 turtle (*Lepidochelys kempii*), which is solely a resident of the Gulf of Mexico and the
124 northeastern coast of North America (Wibbels and Bevan, 2019).

125 Marine mammals such as odontocetes and mysticetes (families: Phocoenidae,
126 Balaenopteridae, Delphinidae, Iniidae/Pontoporiidae), Sirenian and Pinniped (families:
127 Dugongidae, Phocidae and Otariidae) were also included in the review. This group is, as
128 a whole, widely distributed, as there are species that inhabit all the oceans in the world
129 (including those in the Arctic and Antarctic), as in the case of the humpback whale
130 (*Megaptera novaeangliae*) (Cooke, 2018). However, the majority of species are
131 concentrated in the Atlantic, Pacific and Indian oceans, as is the case of the common
132 dolphin (*Delphinus delphis*) (Braulik et al. 2021) and bottlenose dolphin (*Tursiops*
133 *truncatus*) (Wells et al. 2019).

134 The seabird species included in our review belong to families Diomedidae,
135 Procellariidae, Spheniscidae, Anatidae, Phalacrocoracidae, Laridae, Sulidae,
136 Stercorariidae and Oceanitidae. These birds are distributed across all continents, oceans
137 and seas worldwide. Notably, the Mediterranean Sea serves as a prominent breeding
138 and feeding area for several endangered species, such as the Balearic shearwater

139 (*Puffinus mauretanicus*) (BirdLife International. 2018a) and the Yelkouan shearwater
140 (*Puffinus yelkouan*) (BirdLife International. 2018b).

141 The review also encompassed elasmobranchs, which include sharks and rays from
142 diverse families including Sphyrnidae, Triakidae, Carcharhinidae, Myliobatidae,
143 Squalidae, Rajidae, Somniosidae, Rhinobatidae, Brachaeluridae, Lamnidae,
144 Torpedinidae, Scyliorhinidae, Alopiidae, Pseudocarcharidae, Dasyatidae, Mobulidae,
145 Glaucostegidae and Pentanchidae. Many of the species within this group are commonly
146 found in coastal areas and inhabit the oceanic zone delimited by both tropics (Cancer
147 and Capricorn). However, there are exceptions, such as the Greenland shark (*Somniosus*
148 *microcephalus*), which inhabits the Arctic Circle (Kulka et al. 2020).

149 Overall, most of the sea turtle, marine mammal, seabird and elasmobranch species
150 included in the study are classified by the UICN Red List of Threatened Species (UICN,
151 2023) as being Least concern (LC), Near threatened (NT), Vulnerable (VU), Endangered
152 (EN), Critically Endangered (CR) and Data Deficient (DD) (Methods, Table 2 SI).

153 **2.3. Statistical analysis**

154 In order to verify the nature of the data obtained and compiled (Table 1 SI), the normality
155 and homoscedasticity of the data were analyzed by employing R commander (version
156 2.7-0) using the Shapiro-Wilk test and Bartlett test, respectively. The data did not meet
157 these criteria in any of the cases, and we therefore employed a non-parametric Kruskal-
158 Wallis's test. This test was used to examine whether there was a relationship (at the
159 global level) between the percentage of reduction in by-catch and fishing areas, along
160 with the percentage of reduction in by-catch per species group. In both cases, statistical
161 significance was considered when $p < 0.01$. In order to carry out an individual analysis of

162 the relationship between the efficiency of mitigation measures among the fisheries
163 involved by group, a Goodness-of-Fit G-test was carried out by employing the R Studio
164 program (version 4.0.2), using an R Script (Mangiafico, 2015) and the “DescTools” (Andri
165 et al. 2021) and “RVAideMemoire” packages (Hervé, 2023).

166 **3. RESULTS AND DISCUSSION**

167 **3.1. Overview of published literature**

168 Of a total of 389 studies, 316 were excluded on the basis of specific criteria (Figure 1 SI).

169 A total of 73 papers were eventually selected in order to analyze by-catch by year and
170 fishery area. These are shown in Table 1 SI as follows: sea turtles in green, marine
171 mammals in blue, seabirds in orange and elasmobranchs in purple. However, when
172 attempting to assess the number of trials of fishing gear, it was possible to find only 31
173 studies that provided information on the number of trials (8140). Furthermore, studies
174 related to proofs of concept or carried out in laboratories were not taken into account.
175 The number of papers published each year that were found thanks to the literature
176 search are shown in (Table 3 SI), with a maximum of 9 in 2018 and 2020. In contrast, only
177 3 studies were found in 2022 and 2 in 2010.

178 The number of studies assessing by-catch by continent is shown in Figure 2 SI. The
179 continent for which most studies had been conducted was America, with 28 studies,
180 followed by Europe with 22. Within America, the United States carried out the largest
181 number of studies, with 7 in total. In Europe, studies were conducted by 14 different
182 countries, unlike that which occurred with Oceania, where all the studies were carried
183 out in Australia. This figure also shows that there is a lack of studies in developing nations
184 or locations in which small-scale fisheries (SSFs) are widespread, as is the case of Africa,

185 where only 8 studies have been conducted. With regard to the number of trials carried
186 out with each type of fishing gear, gillnets dominated the count with a total of 7636 trials,
187 distantly followed by set net with 273 trials (Figure 3 SI).

188 **3.2. Worldwide mitigation measures**

189 Several mitigation measures with which to minimize the by-catch of these megafauna
190 species have been tested in different fisheries around the world. A summary of the
191 mitigation measures, categorized by the sensory system involved for each group, is
192 provided in Table 1. In the case of sea turtles, the most commonly used olfactory
193 measure involved using an alternative bait type. With regard to visual measures, LED
194 lights emerged as a promising alternative, while in the case of physical measures, various
195 escape options such as TEDs were frequently employed. In the case of odontocetes,
196 mysticetes and other marine mammals such as the harbour seal (*Phoca vitulina*) and the
197 Californian sea lion (*Zalophus californianus*), the measures most frequently used were
198 pingers, a type of Acoustic Deterrent Devices (ADDs). With regard to seabirds, the
199 mitigation measure that predominated was the tori line, a type of seabird scaring device
200 (Gilman et al. 2021) consisting of a line towed from a high point at the aft of a vessel,
201 from which several streamers are attached to scare seabirds and prevent their access to
202 the critical area where baited hooks sink (Domingo et al. 2017). The main mitigation
203 measure used for elasmobranchs was electrosensory devices. However, there was no
204 single predominant measure, with SMART hooks (Grant et al. 2018; O'Connell et al.
205 2014), rare earth (Porsmoguer et al. 2015; Westlake et al. 2018), ferrite magnets
206 (Richards et al. 2018) and Electropositive Metal (EPM) Alloy (Godin et al. 2013;
207 Hutchinson et al. 2012) all being employed. However, it seems that tori lines have

208 obtained good results for some of the species in this group (Jiménez, Forselledo, et al.
209 2019).

210 **3.3. Mitigation measures by group**

211 Overall, it was found that, globally, there were significant differences among the
212 percentage of reduction in by-catch per fishery area ($\chi=36.33$, $df=19$, $p<0.01$, $n=20$,
213 Kruskal-Wallis's test) (Figure 1) and in the percentage of success in by-catch per group
214 ($\chi=14.67$, $df=3$, $p<0.01$, $n=4$, Kruskal-Wallis's test) (Figure 2A).

215 However, no significant differences were found as regards the percentage of reduction
216 in by-catch resulting from different mitigation types ($\chi=10.17$, $df=4$, $p=0.04$, $n=6$, Kruskal-
217 Wallis's test) (Figure 2B).

218 **3.3.1. Sea turtles**

219 In the case of sea turtles, there was no significant difference among the percentage of
220 reduction in by-catch according to the fishery area ($G=9.15$, X-squared $df=10$, $p=0.52$,
221 $n=11$, G-test). Notably, the US Mid-Atlantic bottom trawl gear (Warden, 2011) and the
222 industrial trawling fishery of Gabon (Casale et al. 2017) attained the highest values of
223 reduction in by-catch per individual (mean \pm SD, $0.97 \pm NA$ and $1.0 \pm NA$, respectively).
224 Similarly, no significant differences were observed among sea turtle species ($G=7.46$, X-
225 squared $df=4$, $p=0.11$, $n=24$, G-test). However, the species with the highest percentages
226 of reduction in by-catch reduction were the loggerhead turtle (*Caretta caretta*) ($n=112$)
227 and the olive ridley turtle (*Lepidochelys olivacea*) ($n=114$) (97% and 100%, respectively).
228 This was achieved using physical measures, specifically TEDs (Figure 1).

229 **3.3.1.1. Olfactory**

230 One olfactory measure was investigated by testing the use of alternative bait types to
231 mitigate incidental sea turtle by-catch in longline fisheries. The best results were attained
232 by replacing squid bait with mackerel bait (Figure 3A), reducing by-catch by 88% and 85%
233 for all turtles species (Coelho et al. 2012). In addition, by combining the effects of the
234 mackerel by exchanging the traditional J-style hooks for two circle hooks (one non-offset
235 and one with 10° offset), it was possible to obtain a 50-59% reduction in the by-catch
236 (n=223) (Coelho et al. 2015).

237 **3.3.1.2. Physical**

238 Tori lines were used in a Uruguayan pelagic longline fishery to reduce the by-catch. A
239 reduction of 18.1% was obtained for the loggerhead turtle (n=83), while one of 73.3%
240 was obtained for the leatherback turtle (*Dermochelys coriacea*) (n=15) (Jiménez et al.
241 2019) (Figure 3A).

242 TEDs have also been tested in numerous trawl fisheries, as occurred in the US Mid-
243 Atlantic, with a reduction in by-catch of 97% for the loggerhead turtle (n=112) (Warden,
244 2011). A reduction in by-catch of 100% was documented for Indian fisheries (Raghu et
245 al. 2016), and a similar reduction of 100% was achieved for four sea turtle species
246 (n=131) in Gabonese fisheries (Casale et al. 2017) (Figure 1).

247 **3.3.1.3. Visual**

248 Of all the possible visual mitigations methods for sea turtles, those that have been
249 studied most are LED lights. However, there are also others, such as chemical lightsticks,
250 physical models (predator cut-outs) and buoyless nets (Figure 3A).

251 The implementation of 500 nm green LEDs in an Indonesian small-scale coastal gillnet
252 fishery led to a reduction in multi-species sea turtle by-catch of 61.4% (n=10), and
253 specifically 59.5% of green turtle (*Chelonia mydas*) (n=14) (Gautama et al. 2022). This
254 same measure was also used in a Mexican gillnet fishery, obtaining a reduction of 50%
255 for the loggerhead turtle (n=17) (Senko et al., 2022) and 59% (n=85), 63.9% (n=125) and
256 48.8% (n=41) for green turtles (Wang et al. 2010; Ortiz et al. 2016; Kakai, 2019). Chemical
257 lightsticks, meanwhile, obtained a reduction of 40% for the green turtle (n=85) (Wang et
258 al. 2010). The implementation of 100-400 nm LED lights in a small-scale gillnet fishery
259 led to a decrease in green turtle by-catch of 93% (n=13) (Darquea et al. 2020). In Italian
260 and Mexican fisheries, the use of LEDs led to a reduction in loggerhead turtle by-catch
261 of 100% (n=18) (Lucchetti et al. 2019; Virgili et al. 2018), and a reduction in green turtle
262 by-catch of 39.7% (n=209) when using UV net illumination (Wang et al. 2013). Although
263 TEDs have attained the best results, more trials on LEDs should be included, as they seem
264 to be a good alternative by which to reduce the by-catch of these species.

265 **3.3.2. Marine mammals**

266 In the case of marine mammals, there were no significant differences among the
267 percentages of reduction in by-catch according to fishery areas (G=5.00, X-squared df=5,
268 p=0.41, n=6, G-test). Notably, the highest values of by-catch reduction per individual
269 were attained by the small set net Japanese fishery (Amano et al. 2017) and Norwegian
270 commercial fisheries (Moan and Bjørge, 2021) (mean \pm SD, 10.0 \pm NA and 13.0 \pm 9.8,
271 respectively). Similarly, no significant differences were found among marine mammal
272 species (G=0.57, X-squared df=9, p=0.99, n=11, G-test). However, the narrow-ridged
273 finless porpoise (*Neophocaena asiaeorientalis*) (n=10) and harbour porpoise (*Phocoena*
274 *phocoena*) (n=20) were the species with the highest percentage of reduction in by-catch

275 (100% and 96.9%, respectively) when using acoustic measures, specifically pingers
276 (Figure 1).

277 **3.3.2.1. Acoustic**

278 Sensory technologies, specifically acoustic reflectors and pingers, were designed in the
279 late 1970s and 1980s to deter marine mammals in gillnet fisheries (Dawson, 1991).

280 The AQUAmark 100 pinger, which operates at between 20 and 160kHz, achieved a
281 reduction in narrow-ridged finless porpoise by-catch of 100% (n=10) (Amano et al. 2017)
282 (Figure 1). The long-term effectiveness of the Dukane Netmark 1000 Pinger, which
283 operates at 12-100 kHz harmonics, was assessed in a gillnet fishery (Carretta and Barlow,
284 2011). However, only the by-catch of two species decreased, specifically the common
285 dolphin by 47.4% and the Northern elephant seal (*Mirounga angustirostris*) by 80.8%
286 (n=164). When used in a driftnet fishery, the same pinger achieved reductions of
287 between 18.2% and 100%, depending on the species (Mangel et al. 2013).

288 The evaluation of two pingers in gillnet fisheries, i.e., the Banana pinger by Fishtek
289 Marine Industries (operating at 50-120 kHz, 154 dB) and the Dolphin pinger by Future
290 Oceans (operating at 70 kHz, 132 dB), attained positive results, with reductions of
291 96.90% and 33.50% (Moan and Bjørge, 2021) (Figure 1). Furthermore, the
292 implementation of these devices did not have a major negative impact on their daily
293 fishing operations and contributed to the reduction in marine mammal by-catch.

294 **3.3.2.2. Physical**

295 Berninsone et al. (2020) replaced gillnets with longlines in order to minimize the by-catch
296 of franciscana (*Pontoporia blainvillei*), which it is considered the most threatened
297 cetacean in the South Western Atlantic (Negri et al. 2012; Bordino and Albareda, 2004).

298 This study decreased the by-catch of this species by 90% (n=85) (Figure 3B), thus showing
299 that this method is an excellent alternative. However, only one study was carried out,
300 which makes it difficult to generalize its effectiveness to other areas.

301 **3.3.2.3. Visual**

302 Unlike that which occurs with sea turtles, visual mitigation measures are not commonly
303 used to prevent marine mammal by-catch. Tori lines and Bird Line Weighting (BLW) have
304 been tested in Uruguayan pelagic longline fisheries with no significant results (Jiménez
305 et al. 2019). In order to evaluate the real effect of these measures on marine mammals,
306 further research is consequently necessary.

307 **3.3.2.4. Echolocation reflection**

308 Mysticetes and odontocetes use echolocation, which helps them to determine the
309 location of objects in the sea. A mitigation measure consisting of adding acrylic glass
310 spheres to a gillnet has consequently been developed in order to reduce the by-catch of
311 the harbour porpoise (n=5), obtaining a reduction of 60% (Kratzer et al. 2021). Another
312 measure tested was the modification of two types of nets, a barium sulfate net and a
313 stiff nylon net (Bordino et al. 2013), but a reduction of only 7.4% (n=54) was obtained
314 (Figure 3B).

315 **3.3.3. Seabirds**

316 With regard to seabirds, the percentage of reduction in by-catch was not significantly
317 different among fishery areas ($G=7.84$, X-squared $df=5$, $p=0.16$, $n=6$, G-test). Similarly, no
318 significant differences were found among seabird species ($G=1.58$, X-squared $df=16$, $p=1$,
319 $n=20$, G-test). Notably, the species with the highest percentage of by-catch reduction
320 were the Atlantic yellow-nosed albatross (*Thalassarche chlororhynchos*) (n=43), the

321 black-browed albatross (*Thalassarche melanophris*) (n=22) and the white-chinned petrel
322 (*Procellaria aequinoctialis*) (n=486) (100%, 100% and 97.7%, respectively) when using
323 visual measures, and specifically tori lines (Figure 1).

324 **3.3.3.1. Olfactory**

325 The main olfactory mitigation measures used to minimize seabird by-catch were offal
326 discard management (Kuepfer et al. 2022; Collins et al. 2021; Rollinson et al. 2017),
327 thawed bait (Collins et al. 2021; Rollinson et al. 2017), blue-dyed bait (Gilman et al.
328 2021), artificial bait (Cortés and González-Solís, 2018) and replacing squid with mackerel
329 as bait (Gonzalez et al. 2012; Li et al. 2012) (**¡Error! No se encuentra el origen de la r**
330 **eferencia.**). These measures were, in certain instances, reinforced with non-sensory
331 methods, such as night setting (Collins et al. 2021; Rollinson et al. 2017), seasonal
332 closures (Collins et al. 2021), hook management (Collins et al. 2021) and the limitation
333 of by-catch rates per year (Rollinson et al. 2017).

334 Despite the variety of existing measures, there is limited information regarding their
335 effectiveness as regards reducing seabird by-catch and their effect on commercial
336 catches. For instance, the evaluation of artificial bait demonstrated a reduction in target
337 catches of 77% when compared to control lines (Cortés and González-Solís, 2018), but
338 sample sizes were not included in the study.

339 **3.3.3.2. Physical**

340 In order to reduce the by-catch rate of seabirds by using physical measures, the increase
341 in the sink rate of baited hooks by reducing the distance between the hook and the
342 weight of the branch lines (65g) was tested in a pelagic longline fishery (Jiménez et al.
343 2019), obtaining a reduction of 42.5%. Others studies propose the introduction of BLW

344 as a mitigation measure, such as that by Paterson et al. (2019), which was carried out in
345 a demersal longline fishery where a reduction in the by-catch was from 90.9% to 100%.

346 **3.3.3.3. Visual**

347 The measures most commonly used in the case of seabirds are those of visual mitigation
348 and include techniques such as LEDs (Bielli et al. 2020; Mangel et al. 2018; Field et al.
349 2019), high contrast panels (Field et al. 2019; Oliveira et al. 2021), buoys with looming
350 eyes (Rouxel et al. 2021), night setting (Cortés and González-Solís, 2018) and tori lines
351 (Cortés and González-Solís, 2018; Gilman et al. 2021) (**iError! No se encuentra el origen de
352 la referencia.**).

353 The implementation of 500 nm LEDs was positive as regards reducing the by-catch of 4
354 species (n=46), with a reduction of 84% (Bielli et al. 2020), while green LEDs led to a
355 reduction of 85.1% (Mangel et al. 2018). However, in the study carried out by Field et al.
356 (2019), the efficacy of two types of 500 nm LEDs (constant green lights and flashing white
357 LED lights) achieved a reduction of only 32.6% (n=43), although the use of high contrast
358 panels reduced the by-catch of species by 50.8% (n=65). New devices such as the
359 “Looming eyes buoy” (LEB) have also emerged, leading to a decrease in the by-catch of
360 seabirds species of 22% (n=5724) (Rouxel et al. 2021) (Figure 3C).

361 Another measure is that of night setting, which has been tested with the artisanal
362 demersal longliners of the Western Mediterranean. Although the sample sizes were
363 limited, the results obtained showed a reduction of 83.3% and 100% (n=19) (Figure 3C),
364 and an increase in sample testing is, therefore, recommended in order to ensure the
365 efficiency of this measure.

366 Finally, the most widely used measure with which to reduce the by-catch of seabirds is
367 that of tori lines, which have been shown to provide a significant reduction in the by-
368 catch of all birds species, from 97.7% to 100% (Domingo et al. 2017; Paterson et al. 2019)
369 (Figure 1).

370 **3.3.4. Elasmobranchs**

371 With regard to elasmobranchs, the percentage of reduction in by-catches was
372 significantly different according to the study site ($G=11.83$, X-squared $df=2$, $p<0.01$, $n=3$,
373 G-test), with a Uruguayan longline fishery attaining the highest values as regards a
374 reduction in by-catch (Jiménez et al. 2019). No significant differences were found among
375 elasmobranch species ($G= 1.60$, X-squared $df=17$, $p=1$, $n=21$, G-test), although the
376 species that attained the highest percentage of reduction in by-catch were the night
377 shark (*Carcharhinus signatus*) ($n=38$) and the smooth hammerhead (*Sphyrna zygaena*)
378 ($n=190$) (89.5% and 86.3%, respectively) when using visual measures such as tori lines
379 (Figure 1).

380 **3.3.4.1. Acoustic**

381 During the study of the effectiveness of pingers at reducing the by-catch of certain
382 species of marine mammals, their effect was also analyzed for elasmobranchs. No
383 significant differences in captures were attained when using Aquamark 100 and 200
384 pingers (Bilgin and Kose, 2018; Mangel et al. 2013).

385 **3.3.4.2. Olfactory**

386 Only one olfactory mitigation measure with which to reduce the by-catch of different
387 species of sharks and rays has been studied over a 13-year period (Coelho et al. 2012):

388 that of replacing squid with mackerel bait. However, the effectiveness of this measure
389 was limited.

390 **3.3.4.3. Physical**

391 Physical measures by which to mitigate the by-catch of elasmobranchs include BLW
392 (Jiménez et al. 2019), Bycatch Reduction Devices (BRDs) (Gupta et al. 2020) and the use
393 of a 'tickler' (Kynoch et al. 2015), i.e., a piece of chain placed in front of the bottom gear
394 of the trawler that is considered effective as regards catching skates and rays that may
395 escape under the net. The inclusion of BLW in a Uruguayan longline fishery led to a
396 reduction in the by-catch of the scalloped hammerhead (*Sphyrna lewini*) (n=2) of 100%,
397 while the figure for the pelagic stingray (*Pteroplatytrygon violacea*) (n=18) was 27.8%
398 (Figure 3D). However, in the case of the "tickler", the number of species captured
399 increased for all species with the exception of the lesser-spotted dogfish (*Scylorhinus*
400 *canicular*) (n=1525), which attained a decrease of 2.3% (Kynoch et al. 2015) (Figure 3D).

401 **3.3.4.4. Visual**

402 The use of LED lights as a mitigation measure has been tested for all four groups (sea
403 turtles, marine mammals, seabirds and elasmobranchs), with an uncertain effect on
404 elasmobranchs (Mangel et al. 2018). However, in a Mexican gillnet fishery, there was a
405 reduction in the elasmobranch by-catch of 95% (Senko et al. 2022). Tori lines have also
406 obtained good results for this group, reducing the by-catch rate for the porbeagle
407 (*Lamna nasus*) (n=34), copper shark (*Carcharhinus brachyurus*) (n = 8), night shark (n=38)
408 and smooth hammerhead (n=190) by 41.2%, 87.5%, 89.5% and 86.3%, respectively
409 (Jiménez et al. 2019) (Figure 1).

410 **3.3.4.5. Echolocation**

411 Rays do not use echolocation, and the by-catch data obtained after the implementation
412 of the acrylic glass spheres by Kratzer et al. (2021) in a Turkish commercial fishery
413 confirm this. More thornback skate (*Raja clavata*) individuals were caught in the
414 modified gillnet (n=97) than in the standard one (n=41).

415 **3.3.4.6. Electrosensory**

416 Sharks have a complex and extensive electrosensory system, which includes the
417 ampullae of Lorenzini located around the snout or rostral area (Kajiura and Holland,
418 2002). The use of SMART hooks in a longline fishery in the Gulf of Maine (USA) led to a
419 reduction in the number of shark species caught, from 25% to 100% (O'Connell et al.
420 2014) (Figure 3D).

421 The last sensory type measure found was the use of hooks made from a neodymium-
422 praseodymium alloy, use by longlines in USA and Ecuador with the scalloped
423 hammerhead (n=52), leading to a reduction of 61.5% (Hutchinson et al. 2012) (Figure
424 3D).

425 **3.4. Limitations of this review**

426 The objective of this literature review was to provide a comprehensive overview of the
427 most effective measures used to date in order to reduce the by-catch of sea turtles,
428 marine mammals, seabirds and elasmobranchs. However, it was difficult to carry out the
429 global standardization of data because many studies were incomplete owing to a lack of
430 sample sizes (Königson et al. 2022; Kuepfer et al. 2022; Godin et al. 2013), the existence
431 of small sizes (O'Connell et al. 2014; Domingo et al. 2017; Jiménez et al. 2019) or the
432 absence of the name of the species being studied (*Diomedea spp.*, *Procellaria spp.*,

433 *Delphinus spp.*, *Globicephala spp.*) (Yokota et al. 2011; Mangel et al. 2013). Moreover, in
434 many cases the effectiveness of different mitigation measures for the species involved
435 was not included (Kynoch et al. 2015; Porsmoguer et al. 2015; Basran et al. 2020), making
436 it difficult to make comparisons among studies.

437 **3.5. Recommendations for future research**

438 A limited number of complete studies on by-catch mitigation measures were found.
439 These included TEDs for sea turtles (Warden, 2011; Casale et al. 2017), pingers for
440 marine mammals (Amano et al. 2017; Moan and Bjørge, 2021) and tori lines for seabirds
441 (Domingo et al. 2017; Paterson et al. 2019). However, the available research on this topic
442 is still lacking in many aspects.

443 There is a need for more studies that quantitatively assess the actual amount of by-catch
444 (Basran et al. 2020; Culik et al. 2015; Westlake et al. 2018). Moreover, most of the results
445 obtained often vary according to geographical areas, species and fishing practices, thus
446 highlighting the importance of conducting further research into effective strategies by
447 which to mitigate by-catch, particularly in regions in which SFFs are prevalent, such as
448 Asia, Africa and South America.

449 It is also essential to establish standardized reporting practices, define study parameters,
450 specify research locations and context, and examine unintended impacts on animal
451 populations so as to attain accurate comparisons (Kynoch et al. 2015; Porsmoguer et al.
452 2015; Grant et al. 2018). Consistency in measurement metrics is crucial, focusing on the
453 number of individuals captured per unit of effort (Senko et al. 2022; Berninsone et al.
454 2020; Gautama et al. 2022). Thorough documentation should encompass specifics such
455 as gear type, study locale, the technology employed, and technical specifications (Senko

456 et al. 2022; Mangel et al. 2013). It is, therefore, also recommended that studies explicitly
457 detail sample and effect sizes (O’Connell et al. 2014; Domingo et al. 2017).

458 Furthermore, we believe that it is necessary to increase exploration into the combination
459 of sensory deterrents in order to reduce by-catch across various taxonomic groups
460 (Coelho et al. 2015; Gilman et al. 2021). Future research should prioritize the use of cost-
461 efficient technologies that are straightforward to implement, as these are more likely to
462 gain the support and compliance of the fishing industry. This will make it possible to work
463 toward preserving the future of many endangered species and reducing the impact of
464 by-catch.

465 **4. REFERENCES**

466 Agardy, T. (2000). Effects of fisheries on marine ecosystems: A conservationist’s
467 perspective. ICES Journal of Marine Science, 57(3), 761–765.
468 <https://doi.org/10.1006/jmsc.2000.0721>

469 Amano, M., Kusumoto, M., Abe, M., Akamatsu, T. (2017). Long-term effectiveness of
470 pingers on a small population of finless porpoises in Japan. Endangered Species
471 Research, 32(1), 35–40. <https://doi.org/10.3354/esr00776>

472 Andri, S., Ken, A., Andreas, A., Nanina, A., Tomas, A., Chandima, A., ..., Ben, B. (2021).
473 DescTools: Tools for descriptive statistics. R package version, 0.99, 43.

474 Basran, C. J., Woelfing, B., Neumann, C., Rasmussen, M. H. (2020). Behavioural
475 Responses of Humpback Whales (*Megaptera novaeangliae*) to Two Acoustic Deterrent
476 Devices in a Northern Feeding Ground off Iceland. Aquatic Mammals, 46(6), 584–602.
477 <https://doi.org/10.1578/AM.46.6.2020.584>

- 478 Berninsone, L. G., Bordino, P., Gnecco, M., Foutel, M., Mackay, A. I., Werner, T. B. (2020).
479 Switching Gillnets to Longlines: An Alternative to Mitigate the Bycatch of Franciscana
480 Dolphins (*Pontoporia blainvillei*) in Argentina. *Frontiers in Marine Science*, 7(August), 1–
481 19. <https://doi.org/10.3389/fmars.2020.00699>
- 482 Bielli, A., Alfaro-Shigueto, J., Doherty, P. D., Godley, B. J., Ortiz, C., Pasara, A., Wang, J. H.,
483 Mangel, J. C. (2020). An illuminating idea to reduce bycatch in the Peruvian small-scale
484 gillnet fishery. *Biological Conservation*, 241(August 2019), 108277.
485 <https://doi.org/10.1016/j.biocon.2019.108277>
- 486 Bilgin, S., Kose, O. (2018). Testing two types of acoustic deterrent devices (pingers) to
487 reduce harbour porpoise, *Phocoena phocoena* (Cetacea: Phocoenidae), by catch in
488 turbot (*Psetta maxima*) set gillnet fishery in the Black Sea, Turkey. *Cahiers de Biologie*
489 *Marine*, 59(5), 473–479. <https://doi.org/10.21411/CBM.A.D5B58D5B>
- 490 BirdLife International (2018a). *Puffinus mauretanicus*. The IUCN Red List of Threatened
491 Species. <https://doi.org/10.2305/IUCN.UK.2018-2.RLTS.T22728432A132658315.en>
- 492 BirdLife International (2018b). *Puffinus yelkouan*. The IUCN Red List of Threatened
493 Species. <https://doi.org/10.2305/IUCN.UK.2018-2.RLTS.T22698230A132637221.en>.
- 494 Bordino, P., Albareda, D. (2004). Incidental mortality of franciscana dolphin *Pontoporia*
495 *blainvillei* in coastal gillnet fisheries in Northern Buenos Aires, Argentina. Paper
496 SC/56/SM11. INT. WHAL. COMMN. MEETING, Sorrento, Italy.
- 497 Bordino, P., Mackay, A., Werner, T., Northridge, S., Read, A. (2013). Franciscana bycatch
498 is not reduced by acoustically reflective or physically stiffened gillnets. *Endangered*
499 *Species Research*, 21(1), 1–12. <https://doi.org/10.3354/esr00503>

- 500 Braulik, G., Jefferson, T. A., Bearzi, G. (2021). *Delphinus delphis*. The IUCN Red List of
501 Threatened Species. [https://doi.org/10.2305/IUCN.UK.2021-](https://doi.org/10.2305/IUCN.UK.2021-2.RLTS.T134817215A199893039.e)
502 [2.RLTS.T134817215A199893039.e](https://doi.org/10.2305/IUCN.UK.2021-2.RLTS.T134817215A199893039.e)
- 503 Carretta, J. V., Barlow, J. (2011). Long-term effectiveness, failure rates, and “dinner bell”
504 properties of acoustic pingers in a gillnet fishery. *Marine Technology Society Journal*,
505 45(5), 7–19. <https://doi.org/10.4031/MTSJ.45.5.3>
- 506 Casale, P., Abitsi, G., Aboro, M. P., Agamboue, P. D., Agbode, L., Allela, N. L., Angueko, D.,
507 Bibang Bi Nguema, J. N., Boussamba, F., Cardiec, F., Chartrain, E., Ciofi, C., Emane, Y. A.,
508 Fay, J. M., Godley, B. J., Kouerey-Oliwiwina, C. K., de Dieu Lewembe, J., Leyoko, D.,
509 MbaAsseko, G., ... Formia, A. (2017). A first estimate of sea turtle bycatch in the industrial
510 trawling fishery of Gabon. *Biodiversity and Conservation*, 26(10), 2421–2433.
511 <https://doi.org/10.1007/s10531-017-1367-z>
- 512 Coelho, R., Santos, M. N., Fernandez-Carvalho, J., Amorim, S. (2012). Effects of hook and
513 bait on sea turtle catches in an equatorial Atlantic pelagic longline fishery. *Bull Marine*
514 *Sci*, 88(3) (<https://doi.org/10.5343/bms.2011.1065>), 449–467.
- 515 Coelho, R., Santos, M. N., Fernandez-Carvalho, J., Amorim, S. (2015). Effects of hook and
516 bait in a tropical northeast Atlantic pelagic longline fishery: Part I-Incidental sea turtle
517 bycatch. *Fisheries Research*, 164(3), 302–311.
518 <https://doi.org/10.1016/j.fishres.2014.11.008>
- 519 Collins, M. A., Hollyman, P. R., Clark, J., Soeffker, M., Yates, O., Phillips, R. A. (2021).
520 Mitigating the impact of longline fisheries on seabirds: Lessons learned from the South
521 Georgia Patagonian toothfish fishery (CCAMLR Subarea 48.3). *Marine Policy*, 131(June),
522 104618. <https://doi.org/10.1016/j.marpol.2021.104618>

- 523 Cooke, J. G. (2018). *Megaptera novaeangliae*. The IUCN Red List of Threatened Species.
524 <https://doi.org/10.2305/IUCN.UK.2018-2.RLTS.T13006A50362794.en>.
- 525 Cortés, V., González-Solís, J. (2018). Seabird bycatch mitigation trials in artisanal
526 demersal longliners of the Western Mediterranean. PLoS ONE, 13(5), 1–21.
527 <https://doi.org/10.1371/journal.pone.0196731>
- 528 Crowder, L. B., Hazen, E. L., Avissar, N., Bjorkland, R., Latanich, C., Ogburn, M. B. (2008).
529 The impacts of fisheries on marine ecosystems and the transition to ecosystem-based
530 management. Annual Review of Ecology, Evolution, and Systematics, 39, 259–278.
531 <https://doi.org/10.1146/annurev.ecolsys.39.110707.173406>
- 532 Culik, B., VonDorrien, C., Müller, V., Conrad, M. (2015). Synthetic communication signals
533 influence wild harbour porpoise (*Phocoena phocoena*) behaviour. Bioacoustics, 24(3),
534 201–221. <https://doi.org/10.1080/09524622.2015.1023848>
- 535 Darquea, J. J., Ortiz-Alvarez, C., Córdova-Zavaleta, F., Medina, R., Bielli, A., Alfaro-
536 Shigueto, J., Mangel, J. C. (2020). Trialing net illumination as a bycatch mitigation
537 measure for sea turtles in a small-scale gillnet fishery in Ecuador. Latin American Journal
538 of Aquatic Research, 48(3), 446–455. [https://doi.org/10.3856/vol48-issue3-fulltext-](https://doi.org/10.3856/vol48-issue3-fulltext-2428)
539 2428
- 540 Dawson, S. M. (1991). Modifying gillnets to reduce entanglement of cetaceans. Marine
541 Mammal Science, 7(3), 274–282. [https://doi.org/https://doi.org/10.1111/j.1748-](https://doi.org/https://doi.org/10.1111/j.1748-7692.1991.tb00102.x)
542 7692.1991.tb00102.x

- 543 Domingo, A., Jiménez, S., Abreu, M., Forselledo, R., Yates, O. (2017). Effectiveness of tori
544 line use to reduce seabird bycatch in pelagic longline fishing. *PLoS ONE*, 12(9), 1–15.
545 <https://doi.org/10.1371/journal.pone.0184465>
- 546 FAO (2022). El estado mundial de la pesca y la acuicultura 2022. Hacia la transformación
547 azul. Roma, FAO, 1–257. <https://doi.org/10.4060/cc0461es>
- 548 Field, R., Crawford, R., Enever, R., Linkowski, T., Martin, G., Morkūnas, J., Morkūnė, R.,
549 Rouxel, Y., Opper, S. (2019). High contrast panels and lights do not reduce bird bycatch in
550 Baltic Sea gillnet fisheries. *Global Ecology and Conservation*, 18.
551 <https://doi.org/10.1016/j.gecco.2019.e00602>
- 552 Gautama, D. A., Susanto, H., Riyanto, M., Wahju, R. I., Osmond, M., Wang, J. H. (2022).
553 Reducing sea turtle bycatch with net illumination in an Indonesian small-scale coastal
554 gillnet fishery. November, 1–12. <https://doi.org/10.3389/fmars.2022.1036158>
- 555 Gilman, E., Chaloupka, M., Ishizaki, A., Carnes, M., Naholowaa, H., Brady, C., Ellgen, S.,
556 Kingma, E. (2021). Tori lines mitigate seabird bycatch in a pelagic longline fishery.
557 *Reviews in Fish Biology and Fisheries*, 31(3), 653–666. [https://doi.org/10.1007/s11160-](https://doi.org/10.1007/s11160-021-09659-7)
558 [021-09659-7](https://doi.org/10.1007/s11160-021-09659-7)
- 559 Godin, A. C., Wimmer, T., Wang, J. H., Worm, B. (2013). No effect from rare-earth metal
560 deterrent on shark bycatch in a commercial pelagic longline trial. *Fisheries Research*,
561 143, 131–135. <https://doi.org/10.1016/j.fishres.2013.01.020>
- 562 Gonzalez, A., Vega, R., Barbieri, M. A., Yanez, E. (2012). Determinación de los factores
563 que inciden en la captura incidental de aves marinas en la flota palangrera pelágica

- 564 chilena. Latin American Journal of Aquatic Research, 40(3), 786–799.
565 <https://doi.org/10.3856/vol40-issue3-fulltext-25>
- 566 Grant, S. M., Sullivan, R., Hedges, K. J. (2018). Greenland shark (*Somniosus*
567 *microcephalus*) feeding behavior on static fishing gear, effect of SMART (Selective
568 Magnetic and Repellent-Treated) hook deterrent technology, and factors influencing
569 entanglement in bottom longlines. PeerJ, 2018(5). <https://doi.org/10.7717/peerj.4751>
- 570 Gupta, T., Booth, H., Arlidge, W., Rao, C., Manoharakrishnan, M., Namboothri, N.,
571 Shanker, K., Milner-Gulland, E. J. (2020). Mitigation of Elasmobranch Bycatch in Trawlers:
572 A Case Study in Indian Fisheries. Frontiers in Marine Science, 7(July), 1–17.
573 <https://doi.org/10.3389/fmars.2020.00571>
- 574 Hervé, M (2023). RVAideMemoire: testing and plotting procedures for bioestistics. R
575 package version 0.9-83–3.
- 576 Hutchinson, M., Wang, J. H., Swimmer, Y., Holland, K., Kohin, S., Dewar, H., Wraith, J.,
577 Vetter, R., Heberer, C., Martinez, J. (2012). The effects of a lanthanide metal alloy on
578 shark catch rates. Fisheries Research, 131–133(December 2011), 45–51.
579 <https://doi.org/10.1016/j.fishres.2012.07.006>
- 580 Jiménez, S., Domingo, A., Forselledo, R., Sullivan, B. J., Yates, O. (2019). Mitigating
581 bycatch of threatened seabirds: the effectiveness of branch line weighting in pelagic
582 longline fisheries. Animal Conservation, 22(4), 376–385.
583 <https://doi.org/10.1111/acv.12472>
- 584 Jiménez, S., Forselledo, R., Domingo, A. (2019). Effects of best practices to reduce seabird
585 bycatch in pelagic longline fisheries on other threatened, protected and bycaught

586 megafauna species. *Biodiversity and Conservation*, 28(13), 3657–3667.
587 <https://doi.org/10.1007/s10531-019-01842-4>

588 Kajiuira, S. M., Holland, K. N. (2002). Electroreception in juvenile scalloped hammerhead
589 and sandbar sharks. *Journal of Experimental Biology*, 205(23), 3609–3621.
590 <https://doi.org/10.1242/jeb.205.23.3609>

591 Kakai, T (2019). Assessing the effectiveness of LED lights for the reduction of sea turtle
592 bycatch in an artisanal gillnet fishery - a case study from the north coast of Kenya. *Marine*
593 *Science*, 18(2), 37–44. <https://doi.org/10.1038/278097a0>

594 Königson, S., Naddafi, R., Hedgårde, M., Pettersson, A., Östman, Ö., Benavente Norrman,
595 E., Amundin, M. (2022). Will harbour porpoises (*Phocoena phocoena*) be deterred by a
596 pinger that cannot be used as a “dinner bell” by seals? *Marine Mammal Science*, 38(2),
597 469–485. <https://doi.org/10.1111/mms.12880>

598 Kratzer, I. M. F., Brooks, M. E., Bilgin, S., Özdemir, S., Kindt-Larsen, L., Larsen, F.,
599 Stepputtis, D. (2021). Using acoustically visible gillnets to reduce bycatch of a small
600 cetacean: first pilot trials in a commercial fishery. *Fisheries Research*, 243(December
601 2020). <https://doi.org/10.1016/j.fishres.2021.106088>

602 Kuepfer, A., Sherley, R. B., Brickle, P., Arkhipkin, A., Votier, S. C. (2022). Strategic
603 discarding reduces seabird numbers and contact rates with trawl fishery gears in the
604 Southwest Atlantic. *Biological Conservation*, 266(January), 109462.
605 <https://doi.org/10.1016/j.biocon.2022.109462>

- 606 Kulka, D. W., Cotton, C. F., Anderson, B., Derrick, D., Herman, K., Dulvy, N. K. (2020).
607 *Somniosus microcephalus*. The IUCN Red List of Threatened Species.
608 <https://doi.org/10.2305/IUCN.UK.2020-3.RLTS.T60213A124452872.en>.
- 609 Kynoch, R. J., Fryer, R. J., Neat, F. C. (2015). A simple technical measure to reduce bycatch
610 and discard of skates and sharks in mixed-species bottom-trawl fisheries. *ICES Journal of*
611 *Marine Science*, 72(6), 1861–1868. <https://doi.org/10.1038/278097a0>
- 612 Li, Y., Browder, J. A., Jiao, Y. (2012). Hook effects on seabird bycatch in the United States
613 Atlantic pelagic longline fishery. *Bulletin of Marine Science*, 88(3), 559–569.
614 <https://doi.org/10.5343/bms.2011.1039>
- 615 Lucas, S., Berggren, P. (2023). A systematic review of sensory deterrents for bycatch
616 mitigation of marine megafauna. *Reviews in Fish Biology and Fisheries*, 33(1), 1–33.
617 <https://doi.org/10.1007/s11160-022-09736-5>
- 618 Lucchetti, A., Bargione, G., Petetta, A., Vasapollo, C., Virgili, M. (2019). Reducing sea
619 turtle bycatch in the mediterranean mixed demersal fisheries. *Frontiers in Marine*
620 *Science*, 6(JUL), 1–12. <https://doi.org/10.3389/fmars.2019.00387>
- 621 Mangel, J. C., Alfaro-Shigueto, J., Witt, M. J., Hodgson, D. J., Godley, B. J. (2013). Using
622 pingers to reduce bycatch of small cetaceans in Peru’s small-scale driftnet fishery. *Oryx*,
623 47(4), 595–606. <https://doi.org/10.1017/S0030605312000658>
- 624 Mangel, J. C., Wang, J., Alfaro-Shigueto, J., Pingo, S., Jimenez, A., Carvalho, F., Swimmer,
625 Y., Godley, B. J. (2018). Illuminating gillnets to save seabirds and the potential for multi-
626 taxa bycatch mitigation. *Royal Society Open Science*, 5(7), 4–7.
627 <https://doi.org/10.1098/rsos.180254>

- 628 Mangiafico, S. S (2015). An R Companion for the Handbook of Biological Statistics,
629 version 1.3.9, revised 2023. rcompanion.org/rcompanion/
- 630 Moan, A., Bjørge, A. (2021). Pinger trials in Norwegian commercial fisheries confirm that
631 pingers reduce harbour porpoise bycatch rates and demonstrate low level of pinger-
632 associated negative impacts on day-to-day fishing operations. IWC Scientific Committee,
633 68, 1–18.
- 634 Negri, M. F., Denuncio, P., Panebianco, M. V., Cappozzo, H. L. (2012). Bycatch of
635 franciscana dolphins *Pontoporia blainvillei* and the dynamic of artisanal fisheries in the
636 species' southernmost area of distribution. Brazilian Journal of Oceanography, 60(2),
637 149–158. <https://doi.org/10.1590/S1679-87592012000200005>
- 638 O'Connell, C. P., He, P., Joyce, J., Stroud, E. M., Rice, P. H. (2014). Effects of the SMART
639 TM (Selective Magnetic and Repellent-Treated) hook on spiny dogfish catch in a longline
640 experiment in the Gulf of Maine. Ocean and Coastal Management, 97, 38–43.
641 <https://doi.org/10.1016/j.ocecoaman.2012.08.002>
- 642 Oliveira, N., Almeida, A. N. A., Alonso, H., Constantino, E., Ferreira, A., Gutiérrez, I.,
643 Santos, A. N. A., Silva, E., Andrade, J. (2021). A contribution to reducing bycatch in a high
644 priority area for seabird conservation in Portugal. Bird Conservation International, 31(4),
645 553–572. <https://doi.org/10.1017/S0959270920000489>
- 646 Ortiz, N., Mangel, J. C., Wang, J., Alfaro-Shigueto, J., Pingo, S., Jimenez, A., Suarez, T.,
647 Swimmer, Y., Carvalho, F., Godley, B. J. (2016). Reducing green turtle bycatch in small-
648 scale fisheries using illuminated gillnets: The cost of saving a sea turtle. Marine Ecology
649 Progress Series, 545, 251–259. <https://doi.org/10.3354/meps11610>

- 650 Paterson, J. R. B., Yates, O., Holtzhausen, H., Reid, T., Shimooshili, K., Yates, S., Sullivan,
651 B. J., Wanless, R. M. (2019). Seabird mortality in the Namibian demersal longline fishery
652 and recommendations for best practice mitigation measures. *Oryx*, 53(2), 300–309.
653 <https://doi.org/10.1017/S0030605317000230>
- 654 Porsmoguer, S. B., Bănaru, D., Boudouresque, C. F., Dekeyser, I., Almarcha, C. (2015).
655 Hooks equipped with magnets can increase catches of blue shark (*Prionace glauca*) by
656 longline fishery. *Fisheries Research*, 172, 345–351.
657 <https://doi.org/10.1016/j.fishres.2015.07.016>
- 658 Raghu, R., Boopendranath, M. R., Vinod, M. (2016). Performance Evaluation of Turtle
659 Excluder Device off Dhamra in Bay of Bengal Energy analysis of fishing systems View
660 project Responsible fishing View project. *Fishery Technology*, 53(10), 183–189.
661 <https://www.researchgate.net/publication/307594951>
- 662 Richards, R. J., Raoult, V., Powter, D. M., Gaston, T. F. (2018). Permanent magnets reduce
663 bycatch of benthic sharks in an ocean trap fishery. *Fisheries Research*, 208(March), 16–
664 21. <https://doi.org/10.1016/j.fishres.2018.07.006>
- 665 Rollinson, D. P., Wanless, R. M., Ryan, P. G. (2017). Patterns and trends in seabird bycatch
666 in the pelagic longline fishery off South Africa. *African Journal of Marine Science*, 39(1),
667 9–25. <https://doi.org/10.2989/1814232X.2017.1303396>
- 668 Rouxel, Y., Crawford, R., Cleasby, I. R., Kibel, P., Owen, E., Volke, V., Schnell, A. K., Opper,
669 S. (2021). Buoys with looming eyes deter seabirds and could potentially reduce seabird
670 bycatch in gillnets. *Royal Society Open Science*, 8(5).
671 <https://doi.org/10.1098/rsos.210225>

- 672 Seidu, I., Brobbey, L. K., Danquah, E., Oppong, S. K., van Beuningen, D., Seidu, M., Dulvy,
673 N. K. (2022). Fishing for survival: Importance of shark fisheries for the livelihoods of
674 coastal communities in Western Ghana. *Fisheries Research*, 246(January 2021), 106157.
675 <https://doi.org/10.1016/j.fishres.2021.106157>
- 676 Senko, J. F., Peckham, S. H., Aguilar-Ramirez, D., Wang, J. H. (2022). Net illumination
677 reduces fisheries bycatch, maintains catch value, and increases operational efficiency.
678 *Current Biology*, 32(4), 911-918.e2. <https://doi.org/10.1016/j.cub.2021.12.050>
- 679 IUCN (2023). IUCN Red List of Threatened Species. <https://www.iucnredlist.org>
- 680 Virgili, M., Vasapollo, C., Lucchetti, A. (2018). Can ultraviolet illumination reduce sea
681 turtle bycatch in Mediterranean set net fisheries? *Fisheries Research*, 199 (May 2017),
682 1–7. <https://doi.org/10.1016/j.fishres.2017.11.012>
- 683 Wang, J., Barkan, J., Fisler, S., Godinez-Reyes, C., Swimmer, Y. (2013). Developing
684 ultraviolet illumination of gillnets as a method to reduce sea turtle bycatch. *Biology*
685 *Letters*, 9(5), 3–6. <https://doi.org/10.1098/rsbl.2013.0383>
- 686 Wang, J. H., Fisler, S., Swimmer, Y. (2010). Developing Visual deterrents to reduce sea
687 turtle bycatch in gill net fisheries. *Marine Ecology Progress Series*, 408, 241–250.
688 <https://doi.org/10.3354/meps08577>
- 689 Warden, M. L. (2011). Modeling loggerhead sea turtle (*Caretta caretta*) interactions with
690 US Mid-Atlantic bottom trawl gear for fish and scallops, 2005-2008. *Biological*
691 *Conservation*, 144(9), 2202–2212. <https://doi.org/10.1016/j.biocon.2011.05.012>

- 692 Wells, R. S., Natoli, A., Braulik, G. (2019). *Tursiops truncatus*. The IUCN Red List of
693 Threatened Species. [https://doi.org/10.2305/IUCN.UK.2019-](https://doi.org/10.2305/IUCN.UK.2019-1.RLTS.T22563A156932432.en)
694 [1.RLTS.T22563A156932432.en](https://doi.org/10.2305/IUCN.UK.2019-1.RLTS.T22563A156932432.en).
- 695 Westlake, E. L., Williams, M., Rawlinson, N. (2018). Behavioural responses of
696 draughtboard sharks (*Cephaloscyllium laticeps*) to rare earth magnets: Implications for
697 shark bycatch management within the Tasmanian southern rock lobster fishery. Fisheries
698 Research, 200(January), 84–92. <https://doi.org/10.1016/j.fishres.2018.01.001>
- 699 Wibbels, T., Bevan, E. (2019). *Lepidochelys kempii*. The IUCN Red List of Threatened
700 Species. <https://doi.org/10.2305/IUCN.UK.2019-2.RLTS.T11533A155057916.en>
- 701 Yokota, K., Minami, H., Kiyota, M. (2011). Effectiveness of tori-lines for further reduction
702 of incidental catch of seabirds in pelagic longline fisheries. Fisheries Science, 77(4), 479–
703 485. <https://doi.org/10.1007/s12562-011-0357-4>

704 **Table 1.** Mitigation measures proposed by different countries depending on the sensory system involved for each
705 megafauna group, with their references (Methods SI).

		SPECIES						Reference			
		Sea turtles	Cetaceans	Other mammals	Seabirds	Elasmobranchs					
MITIGATION TYPE	Acoustic	Acoustic Deterrent Device	Pinger		X	X		X	(1)		
		Acoustic Harassment Device	Seal scarer		X				(2)		
	Olfactory	Alternative bait type	Mackerel bait replacing squid	X				X	X	(3)	
		Offal discard management	---					X		(4)	
	Physical	Structural changes	Changes of fishing nets			X				(5)	
			Hook change	X						(6)	
			Branch Line Weighting	X		X	X	X		(7)	
		Escape option	Bycatch Reduction Device	X							(8)
			Turtle Excluder Device	X					X		(9)
			Bird exclusion device					X			(10)
	Visual	Lights	LED lights	X				X	X	(11)	
			UV-LED	X						(12)	
			Chemical lightsticks	X						(13)	
		Physical model	Predator cut-outs	X	X			X	X	(14)	
			Looming eyes	X				X		(15)	
			High contrast panels					X		(16)	
			Tori lines	X		X	X	X	X	(17)	
		Gear colour	Colour net/hook					X		(18)	
	Bait colour	Blue dye					X		(19)		
	Echolocation	Acoustic reflection device	Net material alteration		X					(20)	
			Acrylic spheres		X			X		(21)	
	Electrosensory	Modified hook	SMART hook					X		(22)	
		Magnet	Neodymium-iron-boron					X		(23)	
			Ferrite					X		(24)	
		Electropositive Metal Alloy	Neodymium-praseodymium					X		(25)	

706

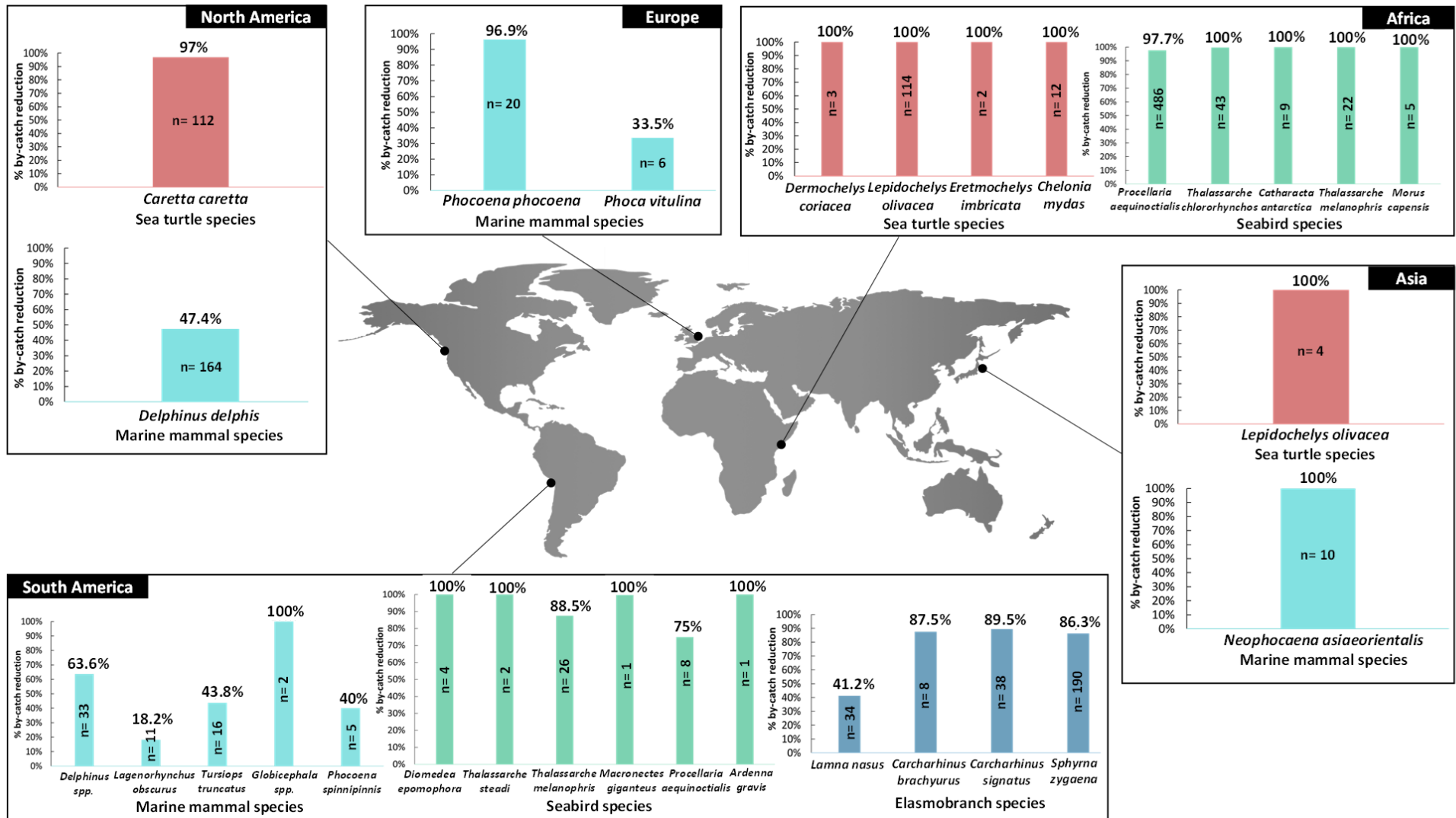


Figure 1. World map showing the percentage of reduction in by-catch per group (sea turtles in red, marine mammals in light blue, seabirds in green and elasmobranchs in dark blue) by continent (America divided into North (1) and South (2) America, Africa (3), Europe (4), Asia (5) and Oceania (6)), obtained by the most frequently used mitigation measure: TEDs for sea turtles, pingers for marine mammals and tori lines for seabirds and elasmobranchs.

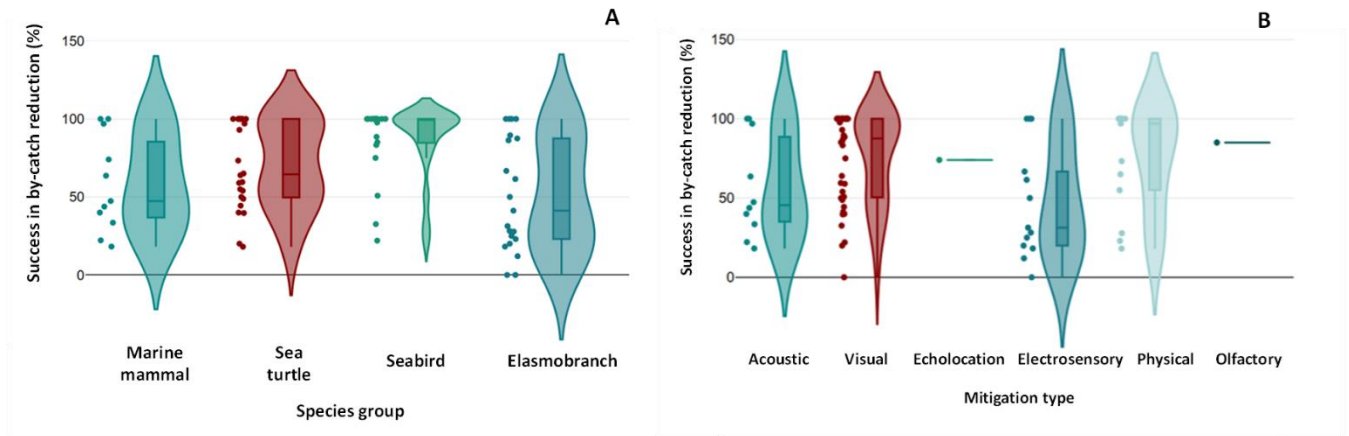


Figure 2. A. Violin plot illustrating the percentage of reduction in by-catch obtained for the species groups (marine turtles, marine mammals, seabirds and elasmobranchs). B. Percentage of reduction in by-catch obtained with the different mitigation types (acoustic, visual, echolocation, electrosensory, physical and olfactory). The horizontal lines inside each box correspond to the mean, while the vertical lines at the ends of each box refer to standard deviation (SD).

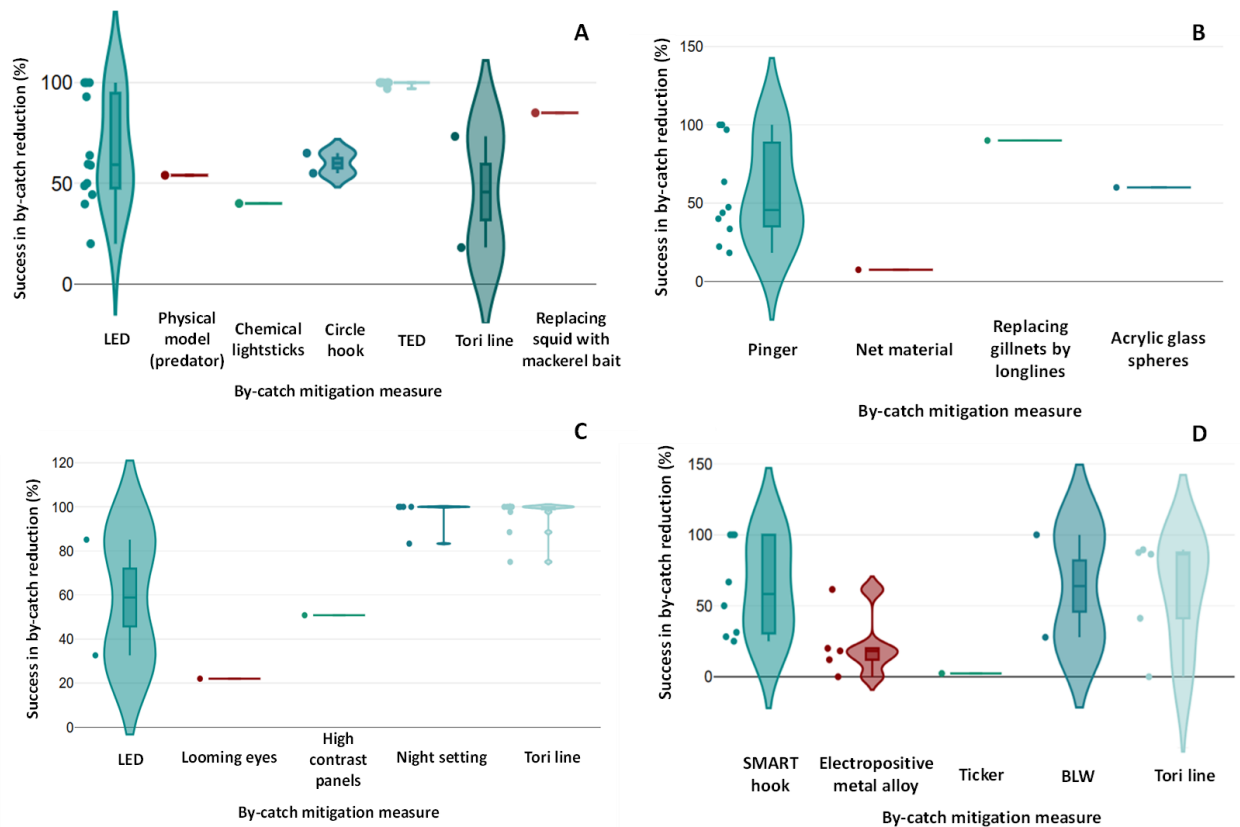


Figure 3. Success in reducing by-catch (%) of A. Sea turtles; B. Marine mammals; C. Seabirds, and D. Elasmobranchs, using different by-catch mitigation measures. The horizontal lines inside each box correspond to the mean, while the vertical lines at the ends of each box refer to SD.